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Radiological Assessment and Potential Health Risks of Tailings from Komu, Southwestern Nigeria

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ABSTRACT

INTRODUCTION

Mining activities often result in the accumulation of various waste materials, commonly referred to as tailings. Tailings usually contain radioactive materials, which have the potential to increase the levels of radioactivity in our environment (Christensen *et al.*, 2014). This increase in radioactivity presents an elevated risk of exposure to ionizing radiation, consequently raising the potential for various radiological health impacts such as cancer, kidney diseases, leukaemia, development of tumours, anaemia, liver complications, bone growth issues, and compromised immune systems (UNSCEAR, 2009; Hossain *et al.*, 2021).

Several studies (Sabo *et al.*, 2018; Ajetunmobi *et al.*, 2019; Usikalu *et al.*, 2019) have highlighted the potential health risks, though they have predominantly concentrated on assessing the levels of radioactivity within the soil in and around mining sites. Despite that, there remains a notable lack of research focusing on the radiological health risks linked with tailings exposure for both miners and residents living in buildings made from such materials, particularly in Southwestern Nigeria. This oversight shows a critical gap in our comprehension of the environmental consequences and possible health hazards associated with tailings.

The escalation of crude mining activities in Nigeria, particularly in communities like Komu town in the Southwest, emphasizes this concern. In Komu, the surge in mining activities and the demand for cost-effective construction materials, often lead to the utilization of tailings. This practice not only exposes miners to outdoor radiation hazards but also exposes residents living inside buildings constructed with these contaminated materials to indoor radiation risks.

This research aims to fill the knowledge gap by assessing the radioactivity levels and evaluating the associated health risks in tailings originating from pegmatite's mining site in Komu, Southwestern Nigeria. By doing so, this study will contribute to the existing pool of knowledge and provide a critical assessment of the radiological threats presented by these tailings. The results are expected to inform effective mitigation strategies and policy formulations, thereby protecting the environment and public health against the adverse impacts of radiation exposure.

MATERIALS AND METHODS Study Area

Komu is situated within Itesiwaju Local Government Area of Oyo State, Southwestern Nigeria. Positioned approximately 170 km northwest of Ibadan, it occupies coordinates of 8°13'60" N latitude and 3°12'0" E longitude. Komu experiences a savanna climate and harbours an estimated population of around 136,772 inhabitants. Geologically, the region is dominated by Precambrian igneous and metamorphic rocks, collectively known as the basement complex. These rocks host a diverse range of mineral deposits, especially granitic pegmatites, which also consist of gemstones, tantalite, tourmaline, mica, quartzite, garnet, magnetite, feldspar, and zircon, among others. Furthermore, these pegmatites are within older rocks such as gneisses, amphibolite, and granites, formed during the Pan-African Orogeny, an event dating back approximately 550 million years ago (Adetunji and Ocan, 2010).

Figure 1: Map showing some sampling points in Komu area of Oyo State.

Samples collection, preparation, and analysis

Thirty (30) tailing samples were randomly collected by digging to a depth of at least 3 cm and carefully packed into labelled polyethene bags. The samples were then taken to the laboratory and subjected to drying to reduce their moisture levels. Afterwards, the samples were crushed and sieved. Following this, spectrometry analysis of the samples was carried out with the aid of a High-purity Germanium (HpGe) detector at the National Institute of Radiation Protection and Research (NIRPR) under the Nigerian Nuclear Regulatory Authority (NNRA), University of Ibadan, Nigeria (Odoh *et al.,* 2019).

Measurement of Activity Concentrations

The activity concentrations of $226Ra$, $232Th$ and $40K$ in the tailing samples were estimated using the formula (Ridha and Hasan, 2016; Alausa *et al.*, 2020):

$$
A = \frac{N}{(E)(y)(m)(t)}\tag{1}
$$

Where *A* is the activity concentration of a specific radionuclide (in Bq/kg), *N* is the net counts of the radionuclide, *E* is the efficiency of the detector at the energy of the radionuclide, *y* is the emission probability of the radionuclide, *m* stands for the mass of the soil sample (kg), and *t* is the counting time (18,000 s). **Radiological Hazard Indices**

Radium Equivalent Activity

The radium equivalent activity is determined using the model (Jibiri *et al.*, 2009 and Shaheen *et al.,* 2023):

 $Ra_{eq} = 0.077 A_K + A_{Ra} + 1.43 A_{Th}$ Where Ra_{eq} is the radium equivalent activity, A_{Ra} is $226Ra$ activity concentration, A_{Th} is $232Th$ activity concentration, and A_K is ⁴⁰K activity concentration.

External Hazard Index

The radiological suitability of the tailing samples for building construction purposes was determined by the external hazard index (Hex) which was evaluated using the model by Yang *et al.*, 2005:

$$
H_{ex} = \frac{A_K}{4810} + \frac{A_R}{370} + \frac{A_{Th}}{259} \le 1
$$
\n(3)

Where A_{Ra} is ²²⁶Ra activity concentration, A_{Th} is ²³²Th activity concentration, and A_K is ⁴⁰K activity concentration.

Internal Hazard Index

The internal exposure to radon and its daughter products was determined by the internal hazard index (Hin), as given by (Tufail *et al.*, 2007):

$$
H_{in} = \frac{A_K}{4810} + \frac{A_R}{185} + \frac{A_{Th}}{259} \le 1
$$
 (4)

Where A_{Ra} is ²²⁶Ra activity concentration, A_{Th} is ²³²Th activity concentration, and A_K is ⁴⁰K activity concentration.

Alpha Representative Index

The Alpha Representative Index (Iα) assesses excess alpha radiation dosage due to radon inhalation by using tailings as building materials. This index was determined using Righi and Bruzzi (2006).

$$
I_{\alpha} = \frac{A_{R\alpha}}{200} \tag{5}
$$

where A_{Ra} is the ²²⁶Ra activity concentration.

Gamma Index

The Gamma Index (Iγ) identifies building materials that may cause health problems due to high quantities of gamma radiation. This assures the safety and well-being of residents in structures built using tailings. The index was estimated using the European Commission's (1999) proposal:

$$
\hat{I}_{\gamma} = \frac{A_K}{3000} + \frac{A_{Ra}}{300} + \frac{A_{Th}}{200}
$$
 (6)
where A_K, A_{Ra}, and A_{Th}, represent ⁴⁰K, ²²⁶Ra, and ²³²Th
activity concentrations, respectively.

RESULTS AND DISCUSSION

Activity Concentrations of Radionuclides

The activity concentrations of ^{226}Ra , ^{232}Th , and ^{40}K radionuclides in the tailing samples from the study area are presented in Table 1. The activity concentration of ²²⁶Ra ranged from 312.07 \pm 8.61 to 851.14 \pm 49.40 Bq/kg, with a mean value of 652.86 ± 33.75 Bq/kg. This mean activity concentration exceeds the global average of 35 Bq/kg for 226 Ra, and values reported by Ademola and Obed (2012) of 39.8 Bq/kg, Suleiman *et al.* (2018) of 49.43 Bq/kg, Osimobi *et al.* (2018) of 33.2 Bq/kg, Focus *et al.* (2021) of 42.59 Bq/kg, and Muhammad *et al.* (2022) of 72.47 Bq/kg.

Similarly, the activity concentration of ²³²Th ranged from 33.31 ± 5.64 to 160.10 ± 15.48 Bq/kg, with a mean value of 116.26 \pm 7.70 Bq/kg, exceeding the world average of 30 Bq/kg. This mean value surpasses reported values by Ademola and Obed (2012) of 17.7 Bq/kg, Ademola *et al.* (2014) of 26.4 Bq/kg, Osimobi *et al.* (2018) of 77.7 Bq/kg, Focus *et al.* (2021) of 35.48 Bq/kg, Muhammad *et al.* (2022) of 86.91 Bq/kg, and Babatunde *et al.* (2023) of 2.85 Bq/kg.

Furthermore, the activity concentration of $40K$ ranged from 96.11 \pm 1.75 to 1184.65 \pm 68.04 Bq/kg, with an average value of 717.53 ± 38.29 Bq/kg. This average value is above the world average of 400 Bq/kg and reported values by Ademola and Obed (2012) of 384.2 Bq/kg, Ademola *et al.* (2014) of 505.1 Bq/kg, Osimobi *et al.* (2018) of 100.7 Bq/kg, Focus *et al.* (2021) of 652.36 Bq/kg, Ma *et al.* (2022) of 533.9 Bq/kg, and Babatunde *et al.* (2023) of 440.45 Bq/kg.

Figure 1: Pie chart depicting the average activity levels of ²²⁶Ra, ²³²Th and ⁴⁰K radionuclides found in the tailings

Radium Equivalent

Table 2 presents the mean, minimum and maximum radium equivalent (Raeq) and hazard indices. The radium equivalent ranged from 426.04 to 1118.38

Bq/kg, with a mean value of 874.36 Bq/kg which is above the ICRP recommended safe threshold of 370 Bq/kg.

(1)						
	Ra _{eq} (Bq/kg)	Hex	Hin	lγ	Iα	
Mean	874.36	2.39	4.15	5.99	3.26	
Minimum	426.04	1.18	2.02	2.99	1.56	
Maximum	118.98	3.06	5.24	7.72	4.26	

Table 2: Radium equivalent (Raeq), external hazard index (Hex), internal hazard index (Hin), gamma index (I_y) , and alpha index (I_a) of the tailing samples

Hazard indices

Figure 2 illustrates the levels of hazard indices considered in this study alongside the permissible limit. The results indicate elevated levels of radiation hazards, as evidenced by the mean external hazard index (Hex) of 2.39, mean internal hazard index (Hin) of 4.15, mean gamma index (Iγ) of 5.99, and mean alpha index (I α) of 3.26. The observation of these values exceeding the threshold of unity (1) raises significant concerns regarding potential radiation hazards.

Figure 2: Comparison of hazard indices of the tailing samples with the permissible limit

CONCLUSION

The concentrations of ^{226}Ra , ^{232}Th and ^{40}K radionuclides in the tailing samples from the study area exceeded the recommended global averages. The heightened levels are likely attributed to geological factors, particularly the presence of Precambrian igneous and metamorphic rocks like granitic pegmatites, as well as intensive mining activities. These activities may disturb the natural equilibrium of radionuclides in the soil, resulting in increased concentrations in the tailings. All the hazard indices considered in the study exceeded the world permissible limit for building materials, indicating increased external and internal exposure risks. This suggests that miners at this site may face an increased risk of radiation exposure and potential health hazards. Based on these findings, the radiological hazards associated with the tailings from Komu's pegmatite mining site are deemed unacceptable, particularly for building purposes. It is therefore recommended that access to this high background radiation area should be restricted with clearly marked signs and physical barriers, and exposure time for workers and the public in this area should be limited.

REFERENCES

Ademola, A. K., Bello, A. K., and Adejumobi, A. C. (2014). Determination of natural radioactivity and hazard in soil samples in and around gold mining area in Itagunmodi, South-western, Nigeria. *Journal of Radiation Research and Applied Sciences, 7(3), 249– 255. https://doi.org/10.1016/j.jrras.2014.06.001.*

Ademola, A., and Obed, R. (2012). Gamma radioactivity levels and their corresponding external exposure of soil samples from tantalite mining areas in Oke-Ogun, South-Western Nigeria. *Radioprotection*, *47*(2), 243–252.

https://doi.org/10.1051/radiopro/2012003.

Adetunji, A., and Ocan, O. O. (2010). Characterization and Mineralization Potentials of Granitic Pegmatites of Komu area, Southwestern Nigeria. *Resource Geology*, *60*(1), 87–97. https://doi.org/10.1111/j.1751- 3928.2010.00116.x.

Ajetunmobi, A., Mustapha, A., Okeyode, I., Gbadebo, A., and Al-Azmi, D. (2019). Assessment of radiological

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safety of abandoned tantalite mining sites in Oke-Ogun, Oyo State, Nigeria. *Radiation Protection and Environment*, 42(1), 40. https://doi.org/10.4103/rpe.rpe_56_18.

Alausa, S. K., Adeyeloja, B., and Odunaike, K. (2020). Radiological Impact Assessment of Farm Soils and Ofada rice (Oryza sativa japonica) from Three Areas in Nigeria. *Baghdad Science Journal*, *17*(3(Suppl.)), 1080. https://doi.org/10.21123/bsj.2020.17.3(suppl.).1080.

Babatunde, A., Oyedokun, O. E., and Olubusola, I. S. (2023). Radiation risk assessment in mining site of Paago, Iseyin Local Govt, Oyo State, Southwestern Nigeria. *Earth Science Malaysia*, *7*(1), 29–35. https://doi.org/10.26480/esmy.01.2023.29.35.

Christensen, D. M., Iddins, C. J., and Sugarman, S. L. (2014). Ionizing Radiation Injuries and Illnesses. *Emergency Medicine Clinics of North America*, *32*(1), 245–265. https://doi.org/10.1016/j.emc.2013.10.002.

European Commission (EC) (1999). Radiological protection principles concerning the natural radioactivity of building materials. *Radiation Protection 112, Directorate General Environment*. Luxembourg: Nuclear Safety and Civil Protection.

Focus, E., Rwiza, M. J., Mohammed, N. K., and Banzi, F. P. (2021). Health Risk Assessment of Trace Elements in Soil for People Living and Working in a Mining Area. *Journal of Environmental and Public Health*, *2021*, 1–10. https://doi.org/10.1155/2021/9976048.

Hossain, S., Anik, A. H., and Alam, M. (2021). Radioactivity and Its Possible Impact on Environment and Human Health: A Review. *International Research Journal of Environmental Sciences*, *10*(4), 24–34.

Jibiri, N. N., Alausa, S. K., and Farai, I. P. (2009). Assessment of external and internal doses due to farming in high background radiation areas in old tin mining localities in Jos-plateau, Nigeria.
Radioprotection, $44(2)$, $139-151$. *Radioprotection,* 44(2), *https://doi.org/10.1051/radiopro/2009001.*

Ma, W., Hu, J., Li, J., Li, J., Wang, P., and Okoli, C. P. (2022). Distribution, source, and health risk assessment of polycyclic aromatic hydrocarbons in the soils from a typical petroleum refinery area in south China. *Environmental Monitoring and Assessment, 194(10). https://doi.org/10.1007/s10661-022-10281-8.*

Muhammad F. R., Gede S. W., and Anung M. (2022). Concentration and Radiological Risk in Tanjung Enim's Coal Mine, South Sumatra Indonesia. *International* *Journal of Cancer Research & Therapy*, *7*(2). https://doi.org/10.33140/ijcrt.07.02.03.

Odoh, C. M., Onudibia, M. E., Iseh, A. J., Ocheje, J. A., Akeredolu, B. J., Ezekiel, Y. A., and Mgbukwu, M. U. (2019). A Comparative Study of NaI(Tl) and HPGe Detectors on Determination of the Activity Concentrations of ⁴⁰K, ²³²Th and ²³⁸U in Soil Samples. *International Journal of Mathematics & Physical Sciences Research.* 7(1): 54-58.

Osimobi, J. C., Avwiri, G. O., and Agbalagba, E. O. (2018). Radiometric and Radiogenic Heat Evaluation of Natural Radioactivity in Soil Around Solid Minerals Mining Environment in South-Eastern Nigeria. *Environmental Processes*, *5*(4), 859–877. https://doi.org/10.1007/s40710-018-0336-1.

Ridha, A. A., and Hasan, H. A. (2016). Cancer Risk Due to the Natural Radioactivity in Cigarette Tobacco. *Detection, 04(03), 54–65. https://doi.org/10.4236/detection. 2016.43008.*

Righi, S. and Bruzzi, L. (2006). Natural radioactivity and radon exhalation in building materials used in Italian dwelling. *Journal of Environmental Radioactivity.* Vol. 88(2): 158 - 170.

Sabo A[,](http://pubs.sciepub.com/jephh/6/2/1/index.html) Sadiq L. S., and Gamba J. (2018). Radiological Assessment of Artisanal Gold Mining Sites in Luku, Niger State, Nigeria. *Journal of Environment Pollution and Human Health*. 2018, 6(2), 45-50. DOI: 10.12691/jephh-6-2-1.

Shaheen, S., Jabbar, A., Ilyas, S. Z., Hussain, M., Dilband, M., Satti, K. H., Shabbir, T., Mehboob, K., and Naseem, A. (2023). Assessment of natural radioactivity levels and potential health risks around coal fired brick kilns of twin cities Rawalpindi and Islamabad, Pakistan. *Arabian Journal of Geosciences*, *16*(7). https://doi.org/10.1007/s12517-023-11507-w.

Suleiman, I., Agu, M., and Onimisi, M. (2018). Evaluation of Naturally Occurring Radionuclide in Soil Samples from Erena Mining Sites in Niger State, Nigeria. *Current Journal of Applied Science and Technology, 27(6), 1–12. [https://doi.org/10.9734/cjast/2018/41562.](https://doi.org/10.9734/cjast/2018/41562)*

Tufail M., Nasim-Akhtar, Sabiha-Javied, and Hamid T. (2007). Natural radioactivity hazards of building bricks fabricated from saline soil of two districts of Pakistan, *Journal of Radiological Protection.* 27, 481-492. https://doi.org/10.1088/0952-4746/27/4/009.

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United Nations Scientific Committee on the Effects of Atomic Radiation (2009). Effects of ionizing radiation, Vol. 1 Annex A. Epidemiological studies of radiation and cancer. *United Nations scientific committee on the effects of atomic radiation (UNSCEAR), United Nations*. http://www.unscear. org/docs/reports.

Usikalu, M. R., Maleka, P. P., Ndlovu, N. B., Zongo, S., Achuka, J. A., and Abodunrin, T. J. (2019). Radiation dose assessment of soil from Ijero-Ekiti, Nigeria. *Cogent Engineering*, *6*(1). https://doi.org/10.1080/23311916.2019.1586271.

Yang, Y., Wu, X., Jiang, Z., Wang, W., Lu, J., Lin, J., Wang, L.M. and Hsia, Y. (2005). Radioactivity concentrations in soils of Xiazhuang granite area, China, *Applied Radiation and Isotope*, Vol. 63, pp.255–259.